

MATRIX MODELS FOR ANALYZING THE POPULATION DYNAMICS OF THE BRAZILIAN MANGROVE CRAB Ucides cordatus L. (DECAPODA: UCIDIDAE)

Modelos matriciais para a análise da dinâmica populacional de *Ucides cordatus L.* (Decapoda: Ucididae)

Marcos de Miranda Leão Leite¹, Dalva Maria Silva-Matos², Raquel Negrão Baldoni³, José Roberto Feitosa-Silva⁴, Cynthia Yuri Ogawa⁵, Luís Ernesto Arruda Bezerra⁶

¹PhD in Ecology and Natural Resources, Graduate Program in Ecology and Natural Resources - PPGERN, Universidade Federal do Ceará - UFC, Avenida Mister Hull s/n Pici, 60.455-760, Fortaleza, CE, Brazil. Faculdade Luciano Feijão – Sobral-CE

²Universidade Federal de São Carlos - UFSCar, Rodovia Washington Luís s/n, 13.565-905, São Carlos, SP, Brazil.

³PhD in Ecology and Natural Resources, Graduate Program in Ecology and Natural Resources - PPGERN, Universidade Federal de São Carlos - UFSCar, Rodovia Washington Luís s/n, 13.565-905, São Carlos, SP, Brazil.

⁴Department of Biology, Universidade Federal do Ceará - UFC, Avenida Mister Hull s/n, Pici, 60.455-760, Fortaleza, CE, Brazil.

⁵Instituto de Ciências Do Mar - LABOMAR, Universidade Federal do Ceará - UFC, Avenida da Abolição 3207, Meireles, 60.165-081, Fortaleza, CE, Brazil.

E-mail author: <u>1975.mirandaleao@gmail.com</u>

ABSTRACT

Ucides cordatus is an important yet declining fishery resource in mangrove ecosystems along the Brazilian coast. This study uses a matrix model to investigate the population dynamics of this species. Specimens were collected monthly using traps from July 2011 and June 2012 in a mangrove area of the Jaguaribe River (Ceará, Northeast Brazil). Population dynamics were analyzed using the size/stage-structured matrix model proposed by Lefkovitch. Carapace width was measured with a stainless steel caliper (accuracy: 0.05 mm). The final sample included 223 female specimens, with an average carapace width of 48.4 mm. Matrix modeling results indicate that the sampled population is expanding (λ =1.30), and the demographic parameter "survival and growth rate" is crucial for maintaining the population.

Keywords: Population dynamics; fisheries management; demography.

Received: 21 May 2024 Approved: 24 May 2024

RESUMO

Ucides cordatus é um recurso pesqueiro dos manguezais da costa brasileira e declínios de suas populações têm sido reportados em diferentes regiões do país. No presente trabalho, foi utilizado um modelo matricial para dar suporte à compreensão da dinâmica de uma população dessa espécie. O estudo foi realizado no manguezal do rio Jaguaribe, estado do Ceará, nordeste do Brasil. Os espécimes foram amostrados mensalmente entre julho de 2011 a junho de 2012. Os caranguejos foram coletados por meio do uso de armadilhas tipo forjo. Para a análise matricial foi utilizado o modelo proposto por Lefkovich, baseado na estrutura de tamanho/estágios. Para isso, os animais tiveram a largura do cefalotórax mensurada por meio de um paquímetro de aço de precisão de 0.05 mm. Foram amostradas 223 fêmeas. A média da largura do cefalotórax foi de 48.4 mm. O resultado da modelagem matricial indicou que a população analisada se encontra em expansão (λ =1.30) e que as taxas vitais relativas à sobrevivência e crescimento são cruciais para a manutenção da população.

Palavras-chave: Dinâmica populacional; manejo pesqueiro; demografia.

INTRODUCTION

Matrix models are useful tools to analyze the population dynamics of plants and animals since they estimate status and identify the demographic parameters/vital rates that most influence the population growth rate (Caswell, 2011). Furthermore, these models provide data for estimating numbers of individuals exploited by extractive activities such as fishing without impairing the viability of populations (Heppel *et al.*, 1996; Miller, 2001; Gross *et al.*, 2002; Freckleton *et al.*, 2003; Rogers-Bennet & Leaf, 2006), which helps define criteria for exploited species management. Criteria for the exploitation of species can be defined using sensitivity and elasticity analyses that identify the critical and priority life cycle phases/vital rates for population abundance and conservation efforts (Begon *et al.*, 2007). The most commonly used models are the Leslie matrix model for age-structured populations and the Lefkovitch matrix model for size or stage-structured populations (Aberg, 1992). Models based on size/stages are ideally used when age is not a good performance predictor of vital rates, such as survival and reproduction (Miller, 2001; Grady & Valiela, 2006). In general, this model takes into account the complexity of species life cycles based on survival rate and permanence in the same stage/size or transition from one stage/size to another.

The mangrove crab Ucides cordatus occurs along the Atlantic coast of the American continent, from Florida (USA) to the state of Santa Catarina (Brazil) (Melo, 1996). This semiterrestrial crab inhabits the intertidal zone of estuarine regions and digs galleries in the soil that can reach depths of 2m (Alcântara-Filho, 1978). It is considered an economically important fishing resource of the Brazilian estuarine regions (Banci et al., 2017; João & Pinheiro, 2018; Machado et al., 2018; Pinheiro et al., 2018; Souza & Pinheiro, 2021; Mota et al., 2023; Pinheiro et al., 2023) and involves a wide supply chain (Ivo & Gesteira, 1999; Glaser & Diele, 2004; Dias-Neto, 2011; Pinheiro et al., 2016; Cortês, 2019). Outside Brazil, U. cordatus is an economically significant fishing resource in Suriname and the Dominican Republic (Nascimento, 1993). This crustacean is manually captured using artisanal methods, although other fishing gear may also be used (Dias-Neto, 2011; Pinheiro et al., 2018; Cortês, 2019). Moreover, it is selectively captured by size and sex and the females are generally returned to the mangrove (Paiva, 1997) together with specimens with cephalothorax less than 6.0 cm wide, which is the minimum catch size established by IBAMA (Ordinance No. 034/03-N, 24.06.2003) (Diele *et al.*, 2005; Leite *et al.*, 2006). In addition to this legal restriction, the market demands larger individuals, preferably males (Botelho et al., 2000; Passos & Di Beneditto, 2005; Fernandes & Carvalho, 2007). A decline in populations of U.

cordatus has been reported in different coastal regions of Brazil (Diele *et al.*, 2005; Legat *et al.*, 2005; Mendonça & Pereira, 2009; Santos *et al.*, 2018; Pinheiro *et al.*, 2023) and the crustacean has been included in the list of species threatened by overexploitation (MMA - Ministério do Meio Ambiente, 2004). However, more recently, its status was changed to "near threatened" due to a 28% decline in the species' population size in Brazilian mangroves (Pinheiro *et al.*, 2016; Mota *et al.*, 2023; Pinheiro *et al.*, 2023).

The rational exploitation of renewable resources is directly related to sustainable development. In this regard, economic and biological factors for the rational use of renewable resources have been considered (Freckleton *et al.*, 2003; Grady & Valiela, 2006). Generally speaking, when the criteria for capturing individuals is larger size, as in the case of *U. cordatus*, fishing should only occur after reproduction to ensure stability of the population (Zhang *et al.*, 2000). The aim of this paper is to understand the population dynamics of the mangrove crab *Ucides cordatus* using matrix modeling and evaluate the population status (decline, equilibrium, or growth) of this species at the study site. Additionally, this paper assesses the current form of exploitation and projects the persistence of stocks based on the following question: "Which life cycle stages are critical for maintaining population abundance?"

MATERIALS AND METHODS

Study site

The study area is located in a mangrove of the Jaguaribe River (4°26'15'' S - 37°48'45'' W), municipality of Aracati, east coast of the state of Ceará, northeast Brazil. The Jaguaribe River Basin is around 633 km long, divided into five sub-basins, and drains a total area of 72,043 km² (Marins *et al.*, 2003). In the estuarine region, the mangrove covers an area of 11.64 km² (Semace, 2006). Regional climate is mild semi-arid tropical with two well-defined seasons, although dry and rainy periods can exceed average rates (Marins *et al.*, 2003). Average temperature ranges between 26 and 28°C (Ipece, 2010). The period of greatest rainfall is from January to May, with a maximum average of 237.8 mm in March. From June to December, the minimum average does not exceed 47.7 mm, while in September, minimum average rainfall is 2.4 mm (Ipece, 2010).

Sampling

Individuals of *U. cordatus* were collected monthly from July 2011 to June 2012 in commercial fishing areas of the community, during low tide periods, using 50 "forjo" crab traps, in each area. This trap is placed in a way that the entrance to the galleries is closed after the crab tries to eat the bait, which consists of red mangrove *Rhizophora mangle* leaves (Carvalho & Igarashi, 2009). During the collections, the traps were placed in the afternoon and collected the next morning, totaling an approximate period of 12 hours. The captured individuals were subjected to thermal shock in a container with ice. They were then transported to the laboratory where sex was recorded by observing morphology of the abdomen (Pinheiro & Fiscarelli, 2001). Cephalothorax width (CW) of the collected crabs was measured using a caliper with 0.05 mm accuracy.

Population dynamics

For matrix analysis, a life table was previously used as a demographic tool. The life table for *U. cordatus* was built according to the method proposed by Murie (1944, *apud* Deevey, 1947), according to the number of individuals killed and age structure of the population. Since chronological marks indicating age were not visible, the individuals were sorted by stage rather

than age. To simplify the model, the different stages of development were based on biological criteria of common characteristics for each of the individuals. In this study, the biological criterion was increased percentage of the species seedling (Diele & Koch, 2010) and size at the beginning of sexual maturation (Leite *et al.*, 2013). A life table was created for the analyzed time interval (2011/2012). The number of stages and their respective intervals were determined as follows: juvenile (20-30 mm), pre-reproductive (30-40 mm), reproductive (40-50 mm), adult I (50-60 mm), adult II (60-70 mm), and senile (70-80 mm).

To calculate the number of survivors (S), total captured individuals should be alive at birth and upon entering the first size class (Murie, 1944 *apud* Deevey, 1947). Thus, it was assumed that the individuals killed by capture would be alive when entering the first size class and, therefore, belong to the same "cohort" that would have been "monitored". Therefore, a sequence of the number of survivors (S) was obtained by backcalculation (Leite *et al.*, 2012). The stages (x) and other parameters of the table were also estimated according to these authors. In addition to survival rates, fertility data is required to construct the matrix (Crouse *et al.*, 1987). Therefore, the model describes the dynamics from the stratum of the population composed of females (Brault & Caswell, 1993).

To construct the life table, the following parameters adapted by Baldoni (2010) were also considered:

Survival rate (sigma-\sigma_{ij}): the survival rate calculated as number of individuals who survived (S) to the subsequent stage divided by the number of initial individuals (N).

Transition rate (gamma-\gamma_{ij}): the transition rate of individuals from one stage (x) to the next (x+1) calculated by considering the number of crabs in each subsequent stage divided by the number of crabs from the previous stage.

The data generated from the life table were analyzed using matrix models. The matrix model used to calculate the population growth rate, sensitivity, and elasticity of each demographic parameter of the transition matrix was based on the model proposed by Lefkovitch (1965). In this model, the matrix elements represent the demographic parameters for each size class or stage rather than age (Caswell, 2001). The transition matrix (A) was constructed as follows:

	P_1	0	F_3	\mathbf{F}_4	F5	F_6
	G_1	\mathbf{P}_2	0	0	0	0
Δ –	0	G_2	\mathbf{P}_3	0	0	0
· • -	0	0	G_3	\mathbf{P}_4	0	0
	0	0	0	G_4	P_5	0
	0	0	0	0	G_5	P_6

In this matrix, **P** represents the probability of surviving and remaining at the same stage; **G** is the probability of survival and growth to the next stage, and **F** represents fertility (Crowder et al., 1994; Caswell, 2001). The probabilities were calculated according to Caswell (2001), as follows:

$$P = \sigma * (1-\gamma)$$
$$G = \sigma * \gamma$$

Fertility rate (F): the fertility rate was calculated using the proposed formula (Grady & Valiela, 2006): Fi = $P_im_i + G_{mi+1}$; where F: fecundity; P= probability of surviving and remaining in the same size class/stage; G= probability of survival and growth to the next size class/stage; m= yield per recruit. This yield was obtained by dividing the number of crabs in the smallest stage at the beginning of the survey by the number of adult females (Gotelli, 2009) from the first stage in which the females are sexually mature, around 45 mm cephalothorax width (Leite et al., 2013).

The population growth rate (lambda - λ) is provided by the eigenvalue of the matrix (Pertierra *et al.*, 1997). Growth rate (λ) is a measure of the balance between survival and reproduction; therefore, when λ >1, the population is growing; when λ <1, the population is

declining, and when λ =1, the population is in equilibrium (Fujiwara & Caswell, 2001). Analyses on the sensitivity of the population growth rate (λ) were performed as a function of changes in the probability that regulates both the transition from stage "I" to any other stage and permanence in "i" (Miller, 2001). Through sensitivity, it is possible to evaluate how the population growth rate responds to changes in demographic rates and identify the parameters to which λ is most dependent (Caswell, 2001). In this study, the sensitivity of (λ) was calculated using the model proposed by Caswell (2001), in which the change in the transition matrix element was calculated as:

$$S_{ij} = \delta \lambda / \lambda a_{ij} = v_i w_i / (v, w)$$

where S_{ij} is the sensitivity of (λ) to changes of the matrix element a_{ij} and v and "w" are the left and right dominant eigenvectors, respectively. Since the probabilities of permanence and transition range from 0 to 1, fecundity is not so limited and it is, therefore, useful to refer to the elasticity of (λ) (Miller, 2001). This is defined as the proportional contribution of a demographic parameter to the population growth rate (de Kroon et al., 2000). According to Caswell (2001), elasticities are calculated as:

 $e_{ij} = a_{ij}/\lambda = \delta\lambda/\delta a_{ij}$

where a_{ij} is a matrix element and λ is the population growth rate.

Since the elasticities are additives, the sum of these elasticities for each stage defines the proportional contribution of the matrix element a_{ij} to global population growth, (λ) (Miller, 2001; Grady & Valiela, 2006). To perform the matrix calculations, the "Popbio" package (Stubben & Milligan, 2007) was used in R software.

RESULTS

A total of 223 crab females were sampled during the study period. Cephalothorax width (CW) ranged from 26.5 to 71.0 mm (mean \pm SD: 48.4 \pm 11.8 mm). The lowest proportions of survivors (σ) in the survey (2011/2012) occurred in the most advanced ontogenetic stages, namely "adult II" and "senile", with 23.7% and 1.3%, respectively. Moreover, the highest transition rates (γ) were observed in the transitions from "juvenile" to "pre-reproductive" (98.6%), "pre-reproductive" to "reproductive" (69.0%), and "reproductive" to "adult I" (67.6%). The lowest transition rate was observed in the passage from "adult II" to "senile" (5.4%). Fertility rates increased from the "reproductive" stage to the "adult II" stage as the "senile" stage decreased in the analyzed annual interval (Table 1).

Table 1 - Life table of the mangrove crab *Ucides cordatus* in the mangrove of Jaguaribe River, northeast Brazil, for the annual interval (2011/2012). D: total deaths; d: number of deaths per stage; σ: survival rate; γ: transition rate; P: probability of survival and permanence at the same stage; G: probability of survival and growth to the next stage; F: fertility rate

Estage	Size Class	d	S	σ	γ	Р	G	F
Juvenile	20 30	3	223	1	0.986	0.014	0.986	
Pre-reproductive	30 40	68	220	0.986	0.690	0.305	0.680	
Reproductive	40 50	49	152	0.681	0.676	0.220	0.460	1.31
Adult	50 60	50	103	0.461	0.514	0.224	0.236	1.47
Adult II	60 70	50	53	0.237	0.054	0.224	0.012	1.83
Senile	70 80	3	3	0.013		0.013		0.96
	D	223						

The transition matrix generated from the vital rates obtained in the life table for 2011/2012 indicated that the population of *U. cordatus* from the mangrove of Jaguaribe River is growing (λ =1.30). Furthermore, in the transition matrix, individuals have a high probability of survival (G) from the "juvenile" stage until reaching sexual maturity. Individuals in the "adult I" and "adult II" categories obtained the highest fertility rates and potentially the highest reproductive values for the population (Table 2).

Juvenile	Pre-reproductive	Reproductive	Adult I	Adult II	Senile
0.014	0	1.31	1.47	1.83	0.96
0.986	0.305	0	0	0	0
0	0.680	0.220	0	0	0
0	0	0.460	0.224	. 0	0
0	0	0	0.236	0.224	0
0	0	0	0	0.012	0.013

Table 2 - Transition matrix obtained for the population of *U. cordatus* in the mangrove of JaguaribeRiver, northeast Brazil, in the interval (2011/2012)

Figure 1 - Simplified arrow diagram of the mangrove crab *Ucides cordatus* in the mangrove of Jaguaribe River, generated from the vital rates obtained in the transition matrix in the analyzed time interval. (P): probability of survival and permanence at the same stage; (G): probability of survival and growth to the next stage; (F): fertility rate



The greatest sensitivity of λ to a change of a matrix element was observed in the transition from the "pre-reproductive" (30-40 mm) to "reproductive" (40-50 mm) stages in the analyzed annual interval (Figure 2).

Figure 2 - Sensitivity of λ to the change in probability of survival and permanence in the same stage (P); probability of survival and growth to the next stage (G); and fertility rate (F)



The matrix elements for vital survival and growth rates from "juveniles" (20-30 mm) to "pre-reproductive" (30-40 mm) and from this stage to "reproductive" (40-50 mm) showed the highest elasticities in the analyzed time interval (Figure 3). Thus, surviving and moving to the next stage is the vital rate with the highest proportional contributions and the rate that best explains the λ value, followed by the fertility rate of reproductive individuals and the probability of an individual surviving and remaining in the same stage.





DISCUSSION

In terms of local dynamic processes, matrix modeling indicated that the growth rate of the *U. cordatus* population was 30% per year (λ =1.30). Populations in equilibrium exhibit a population growth rate of around 1 (Fujiwara & Caswell, 2001). Therefore, the studied population is expanding. This result differs from the findings of Miller (2001), who used matrix models and

identified a decline of a blue crab *Callinectes sapidus* population at a rate of 25% per year. The population of crab species studied by this author is subjected to unsustainable fisheries that catch small individuals and mature, ovigerous females. These demographic categories play an important role in regulating the population dynamics of the species. Thus, this may have influenced the decrease in population growth rate.

The population biology of *U. cordatus* at the same site of the present study and found that males reach sexual maturity when they are 15% larger than the females, size structure indicates relative stability of the population, and the sex ratio is deviated for males. Thus, there was no strong evidence of overexploitation (Leite *et al.*, 2013). According to the authors and the assessment of these indicators, the analyzed population is apparently not suffering the negative effects caused by selective fishing, such as changes in population structure, related to reduced body size, decreased size for sexual maturity, and deviation of sex ratio (Roy *et al.*, 2003; Sato & Goshima, 2006; Fenberg & Roy, 2007). However, Pinheiro *et al.* (2023) observed a drastic reduction in the abundance of *U. cordatus* in the mangroves of Iguape (southeast Brazil) related to anthropogenic impacts which inducing habitat loss. To preserve the species' populations, these authors recommend designating this area as a fishing exclusion zone.

The fishing of *U. cordatus* is selective by sex and size (Botelho *et al.*, 2000; Passos & Di Beneditto, 2005; Fernandes & Carvalho, 2007) and females are generally returned to the mangrove (Paiva, 1997; Diele & Koch, 2010). Therefore, it is considered a potentially sustainable fishery (Diele, 2000). By reducing the fishing pressure on females, the effect of this extrinsic cause of death on the stratum of the population chiefly responsible for regulating population dynamics, through fecundity and recruitment, can be substantially reduced. In this regard, the population growth indicated by modeling may reflect these factors, as opposed to the findings of Miller (2001) for *C. sapidus*.

For *U. cordatus*, the finite rate of population increase (λ) was more sensitive to the transition from "pre-reproductive" to "reproductive" stages, while elasticity analyses revealed that the survival and growth of juveniles up to the "reproductive" stage contributed more proportionally to (λ). According to these results, the main processes that interfere with population dynamics are linked to survival and growth, followed by fecundity. The importance of vital rates for population growth may vary between different taxa. For the species of the Atlantic horseshoe crab *Limulus polyphemus*, the population growth rate was more sensitive to late juvenile survival rates than to other stages (Grady & Valiela, 2006). Therefore, these authors recommend that the capture of individuals before sexual maturity should be avoided.

In contrast, in vertebrate species, (λ) may be more sensitive to changes in the survival of adults (Crouse *et al.*, 1987; Brault & Caswell, 1993; Heppell *et al.*, 1996), to which protection and management should be directed. In the present study, the transition of individuals from the juvenile to adult stages is critical for population maintenance; therefore, efforts should focus on protection to allow individuals to grow, reach sexual maturity, and reproduce. This process that supports the viability and persistence of the population is not under pressure due to the exploitation characteristics of *U. cordatus*. In this regard, the results of matrix modeling revealed that individuals have a high probability of surviving, reaching sexual maturity, and contributing to population replacement. According to Miller (2001), matrix models have some weaknesses despite their usefulness in the study of population dynamics. Specifically in his study, the life story of the blue crab *Callinectes sapidus* involves an inherent spatial component related to larval dispersion and this spatial variability is not incorporated into the model.

To guarantee the conservation of this fishery resource, the current form of exploitation based on selecting catch by size and sex can be maintained. Additionally, environmental agencies must increase enforcement efforts to protected the species during its reproductive season and prevent the extraction of ovigerous females and animals with carapace width of less than 6.0cm. However, factors of this population (sexual dimorphism, sex ratio, size at sexual maturity) should not be altered. Given the economic importance of *U. cordatus* along the Brazilian coast, its populations and exploitation should be monitored to detect population attributes and ensure the necessary management measures. Finally, it is essential to preserve the mangrove areas that serve for larvae settlement, nursery, and refuge of juveniles for their survival, growth, and reproduction

(Miller, 2001; Diele *et al.*, 2005, Schimidt & Diele, 2009; Pinheiro *et al.*, 2023).

ACKNOWLEDGMENTS

The first author would like to thank CAPES (*Coordenação de Aperfeiçoamento de Pessoal de Nível Superior*) for the doctoral grant. We also thank Ronaldo Gonzaga da Silva for his help collecting the animals and all his support in the fieldwork. All samplings were carried out in accordance with federal laws (ICMBio – 20240-1).

REFERENCES

Aberg, P. Size-based demography of the seaweed *Ascophyllum nodosum* in stochastic environments. *Ecology*, v.73, p.1488-1501, 1992.

Alcântara-Filho, P. Contribuição ao estudo da biologia do caranguejo-uçá *Ucides cordatus cordatus* (Linnaeus, 1763) (Crustacea, Decapoda, Brachyura), no manguezal do rio Ceará (Brasil). *Arquivos de Ciências do Mar*, v.18, p.1-41, 1978.

Baldoni, R.N. *Dinâmica de população de* Tapirira guianensis *AUBL. (Anacardiaceae), em áreas de restinga e cerradão do Estado de São Paulo.* Master's Dissertation, Federal University of São Carlos (UFSCar), São Carlos, 2010.

Banci, K.R.S; Mori, G.M. ; Oliveira, M.A.; Paganelli, F.L.; Pereira, M.R.; Pinheiro, M.A.A. Can environmental pollution by metals change genetic diversity? *Ucides cordatus* (Linnaeus, 1763) as a study case in Southeastern Brazilian mangroves. *Marine Pollution Bulletin*, v. 116, p. 440-447, 2017. Begon, M.; Townsend, C.R.; Harper, J.L. (2007). *Ecologia: de indivíduos a ecossistemas*. Oxford: Blackwell Publishing, 759p., 2007.

Botelho, E.R.O., Santos, M.C.R. & Pontes, A.C.P. Algumas considerações sobre a redinha na captura do caranguejo-uçá *Ucides cordatus* (Linnaeus, 1763) no litoral sul de Pernambuco-Brasil. *Boletim Técnico Científico CEPENE*, v.8, p.55-71, 2000.

Brault, S. & Caswell, H. Pod-specific demography of killer whales (*Orcinus orca*). *Ecology*, v.74, p.1444-1454, 1993.

Carvalho, H.R.L. & Igarashi, M.A. A utilização do forjo na captura do caranguejo-uçá (*Ucides cordatus*) na comunidade de Tapebas em Fortaleza-CE. *Biotemas*, v.22, p.69-74, 2009.

Caswell, H. *Matrix Population Models: Construction, Analysis and Interpretation*. Sunderland: Sinauer Associates, 722p., 2001.

Cortês, L.H.O. *Sustentabilidade da extração do caranguejo-uçá* Ucides cordatus (*Linnaeus, 1763*) *no norte do Rio de Janeiro*. Tese de Doutorado em Ecologia e Recursos Naturais, Universidade Estadual do Norte Fluminense Darcy Ribeiro, Campos do Goytacazes, Rio de Janeiro, 2019.

Crouse, D.T., Crowder, L.B. & Caswell, H. A stage-based population model for loggerhead sea turtles and implications for conservation. *Ecology*, v.68, p.1412-1423, 1987.

Crowder, L.B., Crouse, D.T.; Heppell, S.S. & Martin T.H. Predicting the impact of turtle excluder devices on loggerhead sea turtle populations. *Ecological Applications*, v.4, p.437-445, 1994.

de Kroon, H., Groenendael, J.V. & Ehrlen J. Elasticities: a review of methods and model limitations. *Ecology*, v.81, p.607-618, 2000.

Deevey, E.S. (1947). Life tables for natural populations of animals. *Quarterly Review of Biology*, v.22, p.283-314, 1947.

Dias-Neto, Proposta de plano nacional de gestão para o uso sustentável do caranguejo-uçá, guaiamum e do siri-azul. Brasília: Ibama, 156p., 2011.

Diele, K. & Koch, V. Growth and mortality of the exploited mangrove crab *Ucides cordatus* (Ucididae) in N-Brazil. *Journal of Experimental Marine Biology and Ecology*, v.395, p.171-180, 2010

Diele, K. 2000. *Life history and population structure of the exploited mangrove crab* Ucides cordatus cordatus (*L.*) (*Decapoda:Brachyura*) in the Caeté estuary, North Brazil. Tese de Doutorado, Zentrum fur Marine Tropenokologie – ZMT, Bremen.

Diele, K., Koch, V. & Saint-Paul, U. Population structure, catch composition and CPUE of artisanally harvest mangrove crab *Ucides cordatus* (Ocypodidae) in the Caeté estuary, North Brazil: Indications for overfishing? *Aquatic Living Resources*, v.18, p.169-178, 2005.

Fenberg, P.B. & Roy, K. Ecological and evolutionary consequences of size selective harvesting: how much do we know? *Molecular Ecology*, v.17, p.209-220, 2007.

Fernandes, M.E.B. & Carvalho, M.L. Bioecologia de *Ucides cordatus* Linnaeus, 1763 (Decapoda: Brachyura) na costa do estado do Amapá. *Boletim do Laboratório de Hidrobiologia*, v.20, p.15-22, 2007.

Freckleton, R.P., Matos, D.M.; Bovi, M.L.A. & Watkinson A.R. Predicting the impacts of harvesting using structured population models: the importance of density-dependence and time of harvesting for a tropical palm tree. *Journal of Applied Ecology*, v.40, p.846-858, 2003.

Fujiwara, M. & Caswell, H. Demography of the endangered North Atlantic right whale. *Nature*, v.414, p.537-541, 2001.

Glaser, M. & Diele, K. Asymmetric outcomes: assessing central aspects of the biological, economic and social sustainability of a mangrove crab fishery, *Ucides cordatus* (Ocypodidae) in North Brazil. *Ecological Economics*, v.49, p.361-373, 2004.

Gotelli, N. *Ecologia*. Editora Planta, 288p., Londrina, 2009.

Grady, S.P. & Valiela, I. Stage-structured matrix modeling and suggestions for management of Atlantic horseshoe crab, *Limulus polyphemus*, populations on Cape Cod, Massachusetts. *Estuaries and Coasts*, v.29, p.685-698, 2006.

Gross, M.R., Repka, J.; Robertson, C.T.; Secor, D.H. & Winkle, W.V. Sturgeon conservation: Insights from elasticity analysis. *American Fisheries Society Symposium*, 2002.

Heppel, S.S., Caswell, H. & Crowder, L.B. Life histories and elasticity patterns: perturbation analysis for species with minimal demographic data. *Ecology*, v.81, p.654-665, 2000.

Heppel, S.S., Crowder, L.B. & Crouse, D.T. Models to evaluate headstarting as a management tool for long-lived turtles. *Ecological Applications*, v.6, p.556-565, 1996.

Ipece. *Plano básico municipal – Aracati*. Instituto de Pesquisa e Estratégia Econômica do Ceará. Governo do Estado do Ceará, 2010.

Ivo, C.T.C. & Gesteira, T.V.C. Sinopses sobre a biologia do caranguejo-uçá *Ucides cordatus cordatus* (Linnaeus, 1763), capturados em estuários de sua área de ocorrência. *Boletim Técnico Científico CEPENE*, v.7, p.9-50, 1999.

João, M.C.A.; Pinheiro, M.A.A. Reproductive potential of *Ucides cordatus* (Linnaeus, 1763) (Decapoda: Brachyura: Ocypodidae) from two mangrove areas subject to different levels of contaminants. *Journal of Crustacean Biology*, v. 39, p. 74-81, 2018.

Kritzer, J.P. & Sale, P.F. Metapopulation ecology in the sea: from Levins' model to marine ecology and fisheries science. *Fish and Fisheries*. v.5, p.131-140, 2004.

Lefkovitch, L.P. The study of population growth in organisms grouped by stages. *Biometrics*, v.21, p.437-445, 1965.

Legat, J.F.A., Puchnick, A.L., Castro P.F.; Pereira; A.M.L.; Góes, J.M. & Fernandes-Góes, L.C. Current fishery status of *Ucides cordatus* (Linnaeus, 1763) (Brachyura: Ocypodidae) in Parnaíba delta region, Brazil. *Nauplius*, v.13, p.65-70, 2005.

Leite, M.M.L. Rezende, C.F. & Silva, J.R.F. (2012). Tabela de vida do caranguejo-uçá, *Ucides cordatus* (Linnaeus, 1763) (Decapoda: Ucididae), no manguezal do rio Coreaú, nordeste do Brasil. *Arquivos de Ciências do Mar*, v.45, p.75-81, 2012.

Leite, M.M.L., Fonteles-Filho, A.A., Silva, J.R.F. & Cardoso, N.S. Maturidade reprodutiva funcional do caranguejo-uçá, *Ucides cordatus* (Crustacea: Decapoda), no estuário do rio Coreaú, Camocim, Ceará. *Boletim Técnico Científico CEPENE*, v.14, p.41-49, 2006.

Leite, M.M.L., Rezende, C.F. & Silva, J.R.F. (2013). Population biology of the mangrove crab *Ucides cordatus* (Decapoda: Ucididae) in an estuary from semiarid Northeastern Brazil. *Revista de Biología. Tropical*, v.61,n.4, p.1721-1735, 2013.

Machado, I.C.; Barros, M.R.; Piccolo, N.; Matsunaga, A. M. F.; Pinheiro, M. A. A. The capture of the mangrove crab (*Ucides cordatus*) in the estuarine system of Santos-São Vicente: ethnoecology of the fishermen from vila dos pescadores, Cubatão (SP), Brazil. *Boletim do Instituto de Pesca*, v. 44, p. 257, 2018.

Marins, R.V., Lacerda, L.D.; Abreu, E.V.I. & Dias, F.J.S. Efeitos da açudagem no rio Jaguaribe. *Ciência Hoje*, v.33, p.66-70, 2003.

Melo, G.A. *Manual de identificação dos Brachyura (Caranguejos e Siris) do Litoral Brasileiro*. Editora Plêiade, FAPESP (Fundação de Amparo à Pesquisa do Estado de São Paulo), São Paulo, 604p., 1996.

Mendonça, J.T. & Pereira, A.L.C. (2009). Avaliação das capturas de caranguejo-uçá *Ucides cordatus* no município de Iguape, litoral sul de São Paulo, Brasil. *Boletim do Instituto de Pesca*, v.35, p.169-179, 2009.

Miller, T.J. (2001). Matrix-based modeling of blue crab population dynamics with applications to Chesapeak Bay. *Estuaries,* v.24, p.535-544, 2001.

MMA - Ministério do Meio Ambiente. *Instrução normativa n. 5, de 21 de maio de 2004 – Anexo II: Lista Nacional das espécies de invertebrados aquáticos e peixes sobreexplotadas ou ameaçadas de sobreexplotação*, 2004.

Mota, T.A.; Pinheiro, M.A.A.; Evangelista-Barreto, N.S.; <u>Rocha,S.S</u>. Density and extractive potential of uçá crab, *Ucides cordatus* (Linnaeus, 1763), in mangroves of the -Todos os Santos- Bay, Bahia, Brazil. *Fisheries Research*, v. 265, p. 106733, 2023.

Nascimento, S.A. *Biologia do caranguejo-uçá* (Ucides cordatus). Administração Estadual do Meio Ambiente (ADEMA), Aracaju, 45p., 1993.

Paiva, M.P. *Recursos pesqueiros estuarinos e marinhos do Brasil*. Edições UFC, Fortaleza, 286p., 1997.

Passos, C.A. & Di Beneditto, A.P.M. Captura comercial do caranguejo-uçá *Ucides cordatus* (L., 1763) no manguezal de Gargaú – RJ. *Biotemas*, v.18, p.223-231, 2005.

Pertierra, J.P., Lleonart, J. & Lo, N.C.H. Application of a stage-specific matrix model and lengthcohort based analysis the anchovy fishery in Catalan coastal waters (NW Mediterranean Sea). *Fisheries Research*, v.30, p.127-137, 1997.

Pinheiro, M. & Boos, H. (Org.). *Livro Vermelho dos Crustáceos do Brasil 2010-2014*. Porto Alegre, RS, Sociedade Brasileira de Carcinologia - SBC, 466 p., 2016.

Pinheiro, M.A.A. & A.G. Fiscarelli. *Manual de apoio à fiscalização do caranguejo-uçá* (Ucides cordatus). Unesp/Cepsul/Ibama, Rio Claro, 43p., 2001.

Pinheiro, M.A.A., Santos, L.C.M., Souza, C.A., João, M.C.A., Dias-Neto, J. & Ivo, C.T.C. *Avaliação do caranguejo-uçá Ucides cordatus* (Linnaeus, 1763) (Decapoda: Ucididae). Cap. 33: p. 441-458. In:

Pinheiro, M.A.A.; Sousa, F.V.B.; Perroca, J.F.; Silva, M.M.T.; Sousa, R.L. M.; Mota, T.A.; Rocha, S.S. Advances in population monitoring of the mangrove -uçá-crab (*Ucides cordatus*): reduction of body size variance for better evaluation of population structure and extractive potential. *Marine Biology Research*, v. 19, p. 1-13, 2023.

Pinheiro, M.A.A.; Souza, M.R.; Santos, L.C.M.; Fontes, R. F. C. . Density, abundance and extractive potential of the mangrove crab, *Ucides cordatus* (Linnaeus, 1763) (Brachyura, Ocypodidae): subsidies for fishery management. *Anais da Academia Brasileira de Ciências* (online), v. 90, p. 1381-1395, 2018.

Rogers-Bennet, L. & Leaf, R.T. Elasticity analyses of size-based red and white abalone matrix models: management and conservation. *Ecological Applications*, v.16: p.213-224, 2006.

Roy, K., Collins, A.G., Becker, B.J., Begovic, E. & Engle, J.M. Anthropogenic impacts and historical decline in body size of rocky intertidal gastropods in southern California. *Ecology Letters*, v.6, p.205-211, 2003.

Santos, L.C.M.; Pinheiro, M.A.A.; Dahdouh-Guebas F, Bitencourt M.D. Population status and fishery potential of the mangrove crab, *Ucides cordatus* (Linnaeus, 1763) in North-eastern Brazil. *Journal of the Marine Biological Association of the United Kingdom*. 98(2): 299-309, 2018.

Sato, T. & Goshima, S. Impacts of male-only fishing and sperm limitation in manipulated populations of an unfished crab, *Hapalogaster dentata*. *Marine Ecology Progress Series*, v.313, p.193-204, 2006.

Schmidt, A.J. & Diele, K. First field record of mangrove crab *Ucides cordatus* (Crustacea: Decapoda: Ucididae) recruits co-inhabiting burrows of conspecific crabs. *Zoologia*, v.26, p.792-794, 2009.

Semace. Atlas dos manguezais do nordeste do Brasil: avaliação das áreas de manguezais dos estados do Piauí, Ceará, Rio Grande do Norte, Paraíba e Pernambuco. Fortaleza, Brasil. Semace, 125p., 2006.

Souza, F.V.B.; Pinheiro, M.A.A. Biology, trophic chain, and ethnobiological calendar of the mangrove crab, *Ucides cordatus* (Linnaeus, 1763) (Brachyura, Ocypodidae), according to the perception of catchers in Itanhaém, São Paulo, Brazil. *Nauplius*, v.30, 2021.

Stubben, C.J. & Milligan, B.G. Estimating and analyzing demographic models using the popbio package in R. *Journal of Statistical Software*, v.22, n.11, p.11-23, 2007.

Zhang, X.A.; Chen, L.; Neumann, A.U. The stage-structured predator-prey model and optimal harvesting policy. *Mathematical Biosciences*, v.168,p. 201-210, 2000.